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Managing Groundwater Demand through Surface Water and Reuse

Strategies in an Overexploited Aquifer of Indian Punjab

Abstract

Groundwater sustainability is one of the most critical issues to the State of Punjab, India. This research employed a numerical groundwater flow model (MODFLOW) to simulate flow and groundwater levels in the Sirhind Canal Tract of Punjab between 1998 to 2030. Historical groundwater patterns were calibrated using reported groundwater data from 1998 to 2013 for aquifer parameters viz. hydraulic conductivity and specific yield. Thereafter, the calibrated flow simulated model was validated for the years 2013-2018. Twelve possible strategies, including three irrigation conditions and four pumping scenarios, were postulated to evaluate the performance of groundwater resources through to 2030. During the study, it was found that if current groundwater abstraction continues, there will be a further steep decline of 21.49 m in groundwater level by 2030. Findings also suggest that canal water supplies will be beneficial to reverse groundwater level decline and help to increase the water level by 11% above that in the year 2018. The projected increases in water level will reduce energy demand leading to reduced CO₂ emissions of approximately 966.6 thousand tons by 2030.

Keywords: Groundwater depletion; Wastewater reuse; Sustainable groundwater management Simulation modelling

24 **1. Introduction**

25 Punjab, “The Land of Five Rivers”, occupies 1.53% of the total geographical area of
26 India, with 82% of the State’s entire land being used for agricultural activities (ESOPB 2020,
27 pp.78-79). The cultivation of high yield crop varieties and the increased utilisation of
28 fertilisers and groundwater reserves has resulted in a significant increase in the agricultural
29 production of Punjab during the last few decades, hence establishing the State as the
30 leading supplier of food grains to the country (Aggarwal et al. 2005; Khosa et al. 2012). The
31 role of the State in ensuring national food security, via leveraging an extended mandi supply
32 network system, is prominent as out of the 7,128 thousand ha of the total area under
33 cultivation for food and non-food crops, rice and wheat cover nearly 81% (ESOPB 2020, 78-
34 79, pp.86-89). Approximately 99% of Punjab’s net agricultural area is sown under irrigation,
35 while surface water resources accommodate only 28% of the total irrigation needs (ESOPB
36 2020, p.157). The rest 72% of the required water supplies become available through
37 tubewells (ESOPB 2020, p.157). The increased utilisation of groundwater abstraction
38 structures in the State has increased during the past five decades owing to: (i) the
39 implementation of resource development schemes supported by liberal funding from
40 institutional finance agencies; (ii) the increased availability of electric power and diesel; (iii)
41 the provision of quality seeds and fertilisers; and (iv) direct governmental support via
42 subsidies (Ministry of Water Resources 2017, p.15).

43 However, the economic gains have come at a cost with significant depletion of
44 groundwater resources (Saleth 2004; Shah 2009; Liu et al. 2019). In fact, the current average
45 groundwater extraction rate in Punjab is 165%, with the sustainable threshold being 20

46 billion m³ (Central Ground Water Board 2017, p.13; Srivastava et al. 2015), hence resulting
47 in a cumulative groundwater level decline of 10.8 m over the 20 year period, 1998-2018, with
48 an annual fall rate of 54 cm. Based on the Central Ground Water Board (2017), the aquifers
49 in Punjab are projected to run completely dry within the next 25 years if the current
50 groundwater abstraction rate is maintained.

51 **India ranked top in groundwater abstraction in 2010 because of its tenfold increase**
52 **in groundwater abstraction in the past 50 years, representing twice the annual groundwater**
53 **abstraction of either U.S. or China (Gorelick and Zheng 2015). Whilst irrigation water may be**
54 **drawn from ground and surface water sources (Knapp et al. 2000), the current**
55 **predominance of abstraction at unsustainable levels is alarming.**

56 **Overexploitation of groundwater threatens future availability of the most valuable**
57 **natural resources, increases freshwater resources' pollution, and degrades ecosystems**
58 **(Jha et al. 2006).** The rapid depletion of groundwater resources is a serious national concern
59 regarding agri-food supply networks and food security. Significant threats to water
60 resources around the world have initiated the concern for their sustainable management
61 (Vadiati et al. 2018). The most efficient measure to fight against groundwater pollution is to
62 make a plan for integral prevention (Manos et al. 2010, Esteban and Dinar 2013). In the past
63 decades, many researchers suggested solutions applicable to both upstream and
64 downstream agri-food supply chain operations to promote the restoration of groundwater
65 levels in Punjab, including crop diversification, efficient irrigation options, water
66 conservation practices, groundwater recharge techniques and wastewater reuse (Aggarwal
67 et al. 2004; Aggarwal et al. 2005; Shakoore et al. 2018). From a policy-making perspective,

68 Hira (2009) emphasised shifting subsidies from farming inputs (e.g., subsidised energy to
69 tubewells) to farming outputs (e.g., increase in procurement price or bonus on marketable
70 agricultural produce) to encourage farmers to use irrigation water more efficiently.
71 Furthermore, Kaur et al. (2011) suggested an optimal crop plan that could result in
72 groundwater savings of almost 25% in Punjab. Nevertheless, to the best of our knowledge,
73 very limited research has been conducted on how suggested interventions could impact
74 groundwater behaviour.

75 Groundwater systems are complex in nature entailing a range of constituents, like
76 surface water, geological media containing water (such as aquifers), flow boundaries, water
77 recharge factors, sinks leading to water withdrawals, heterogeneous distribution of
78 subsurface materials, and transient groundwater fluxes (Chang et al. 2017). To this effect,
79 groundwater systems' analysis generally requires sophisticated mathematical
80 tools/models to analyse the underpinning dynamics (Mayer et al. 1998; Patil et al. 2020).
81 Simulation models can help assess the impact of different controls on groundwater
82 recharge (Li et al. 2017; Tubau et al. 2017), provided that the associated models realistically
83 account for all the involved processes. Thereafter, a calibrated and validated model can
84 help identify the critical depletion and recharge zones, hence informing groundwater
85 management strategies.

86 Groundwater models allow stakeholders to simulate and predict changes in water
87 tables and determine water reserves' contamination levels. Models such as MODFLOW,
88 FEFLOW, GMS, MODPATH, RT3D, SEEP2D and SEAM3D have been extensively used to
89 inform groundwater resources management (Kori et al. 2013, Gorelick and Zheng 2015).

90 Indicatively, MODFLOW has been used to inform the management of groundwater reserves
91 via adjusting the current extraction patterns to sustain future water resources (Rejani et al.
92 2008; Shakoor et al. 2018). Nevertheless, modelling studies focusing on the groundwater
93 levels in Punjab, explicitly using up-to-date data, are sparse. In addition, despite the global
94 water withdrawals of more than 65%, irrigated agriculture contributes 44% to the agricultural
95 production, which represents only 18% of the cultivated area (FAO 2014). This further justifies
96 the need to consider the efficiency of water use and sewage water reuse in the applied research
97 methodologies and approaches (Barbosa et al. 2017). The application of treated wastewater in
98 agriculture increases in water scarce areas of Mediterranean countries to augment water supply
99 (Candela et al. 2006). The utilisation of multi-water resources is deemed essential for sustainable
100 development (Wang et al. 2019). For example, farming and the food processing industries
101 generate large volumes of effluents, rich in organic matter, that can be reused as an
102 environmentally sustainable practice for irrigating plantations (Menegassi et al. 2020). To the
103 best of our knowledge, the impact of wastewater augmentation on groundwater levels has
104 not been examined, particularly via investigating a range of alternative irrigation conditions
105 and pumping scenarios. Vij et al. (2021) investigated wastewater paradigms in India during
106 the last three decades and observed a policy shift towards wastewater treatment from
107 *'water resource to meet basic human needs'* to *'water scarcity and beautification of cities'*.
108 However, the authors further realised a gap between policy recommendations and actual
109 on-the-ground interventions.

110 Motivated by the scarcity of studies on groundwater modelling in Indian Punjab, and
111 to further address the evident gap about modelling approaches investigating the respective

112 impact of multi-water resources' management, this research contributes to the extant body
113 of literature on groundwater management via: (i) simulating the groundwater system of the
114 Sirhind Canal Tract, one of the most productive aquifer systems at the Indo-Gangetic basin;
115 and (ii) assessing the groundwater level in the region via evaluating alternative pumping and
116 water management scenarios. Following the suggestion of Miglani et al. (2015), this
117 research applied MODFLOW to simulate and assess the future groundwater behaviour of
118 the Sirhind Canal Tract. The novelty of this research is encapsulated to the following: (i) a range
119 of irrigation conditions was considered, including wastewater augmentation from Sewage
120 Treatment Plants (STPs); and (ii) different pumping scenarios were evaluated, resulting in
121 twelve possible strategies for the sustainable management of groundwater resources by
122 2030. Considering that groundwater behaviour is a dynamic phenomenon and the need to
123 update the respective information regularly, this research used recently gathered data
124 about the region.

125 The remainder of this research is structured as follows. Section 2 provides a research
126 background by summarising recent studies applying MODFLOW for groundwater modelling.
127 Section 3 details the materials and methods pertinent to this research. Thereafter, Section
128 4 discusses the research results and findings. In particular, the impact of alternative
129 management strategies on groundwater levels, along with the associated energy
130 requirements and carbon emissions, is presented. Finally, conclusions, policy-making
131 recommendations, limitations, and future research avenues are explored in Section 5.

132

133 **2. Recent Groundwater Management Studies Using MODFLOW**

134 Groundwater simulation modelling is a valuable analysis approach in regions where
135 groundwater resources' extraction exceeds the natural and induced aquifer recharge rates
136 over long periods, like in the case of the Indian Punjab. To that end, MODFLOW, a three-
137 dimensional finite-difference groundwater flow model developed by the United States
138 Geological Survey, has been extensively used by hydrogeologists for the analysis of
139 groundwater flows due to its easiness, accessibility, and versatility (Kashaigili et al. 2003).
140 Selected studies on groundwater management using MODFLOW are analysed below.

141 Rejani et al. (2008) used the Visual MODFLOW package and tested five pumping
142 strategies for groundwater management against overdraft and seawater intrusion in the
143 Balasore coastal basin in Orissa, India. Furthermore, the application of MODFLOW in the
144 North China Plain showed that 29.2% reduction in irrigation could prevent groundwater
145 depletion, while an additional 10% reduction in pumping could foster the restoration of the
146 groundwater aquifer to the hydrological conditions of 1956 within 74 years (Hu et al. 2010).
147 Singh and Shukla (2016) employed Visual MODFLOW to project groundwater levels and
148 articulate potential groundwater management scenarios for the Sai Gomti interfluvial region
149 in India. Khan et al. (2017) proposed a groundwater policy for the Indus Basin of Pakistan via
150 examining different groundwater modelling scenarios, including: (i) groundwater pumping
151 controls; (ii) improvements to the canal infrastructure; and (iii) precipitation changes. In
152 addition, Sobeih et al. (2017) and Eltarabily et al. (2018) used MODFLOW to model
153 groundwater flows in the western and eastern Nile Delta to assess different scenarios
154 capturing recharge from nearby canals and discharge from wells. Similarly, Shakoor et al.
155 (2018) quantified the future groundwater depletion under three different pumping scenarios

156 for the Punjab province in Pakistan and found that artificial groundwater recharge and
157 surplus canal water supplies could be beneficial to avoid groundwater depletion. Glass et
158 al. (2018) suggested that local overexploitation of groundwater resources in Hanoi, Vietnam,
159 can be reduced by smart relocation of wells from the main depression cones and via the
160 expansion of riverbank filtration.

161 In addition, Bouqdaoui and Aachib (2019) focused on managing and restoring
162 groundwater resources of the Berrechid aquifer in Morocco via examining alternative
163 simplified management scenarios such as artificial recharge and reduction of the water
164 volume pumped for irrigation. Furthermore, Chakraborty et al. (2020) used Visual
165 MODFLOW 2000 for analysing the groundwater level in Purba (East) Midnapur, West Bengal,
166 India. Siva Prasad et al. (2020) investigated the groundwater balance for the Kandivalsa River
167 sub-basin in Andhra Pradesh, India, and examined the impact of two scenarios on
168 groundwater levels, namely: (i) increasing the withdrawal rate and keeping the recharge rate
169 constant; and (ii) increasing the recharge rate by 50% while maintaining a constant
170 withdrawal rate. Table 1 summarises representative studies on groundwater modelling
171 using MODFLOW.

172 **Table 1.** Groundwater modelling using MODFLOW: An overview.

173 [Table 1 about here]

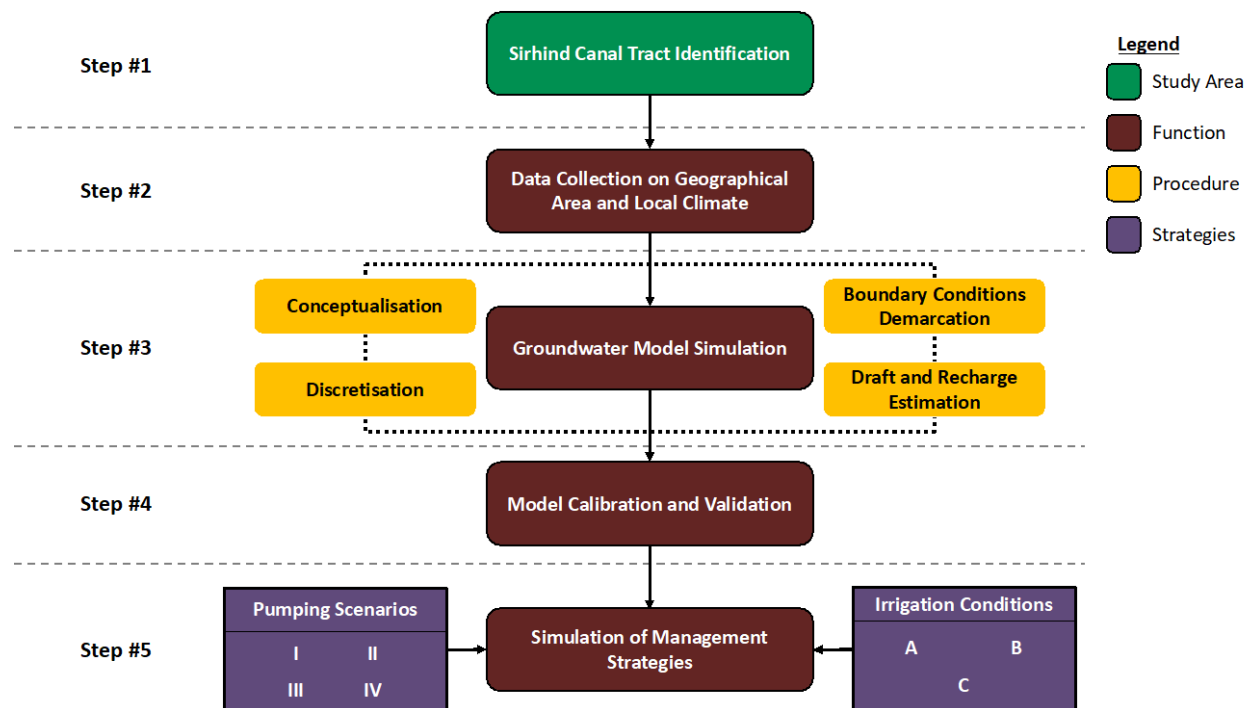
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175 **3. Materials and Methods**

176 This research followed a multi-step methodological approach for developing the
177 elaborated simulation model (Figure 1). Firstly, in Step #1, the Sirhind Canal Tract was

178 identified as the object of scrutiny. Thereafter, in Step #2, the relevant geographic and
 179 climate characteristics were identified while respective data was gathered. MODFLOW-
 180 based simulation model of the Sirhind Canal Tract was conceptualised and discretised in
 181 Step #3. In addition, boundary conditions were set while groundwater draft and recharge
 182 rates were estimated. In Step #4, the model was calibrated and validated to articulate
 183 alternative simulation scenarios. Thereafter, in Step #5, alternative simulation strategies
 184 were investigated. Specifically, four pumping scenarios under three irrigation conditions
 185 were simulated, resulting in twelve alternative strategies that could impact the future
 186 groundwater behaviour. The modelling results were summarised and analysed to suggest
 187 environmentally sustainable policies for the management of the groundwater behaviour in
 188 the Sirhind Canal Tract of the Indian Punjab.

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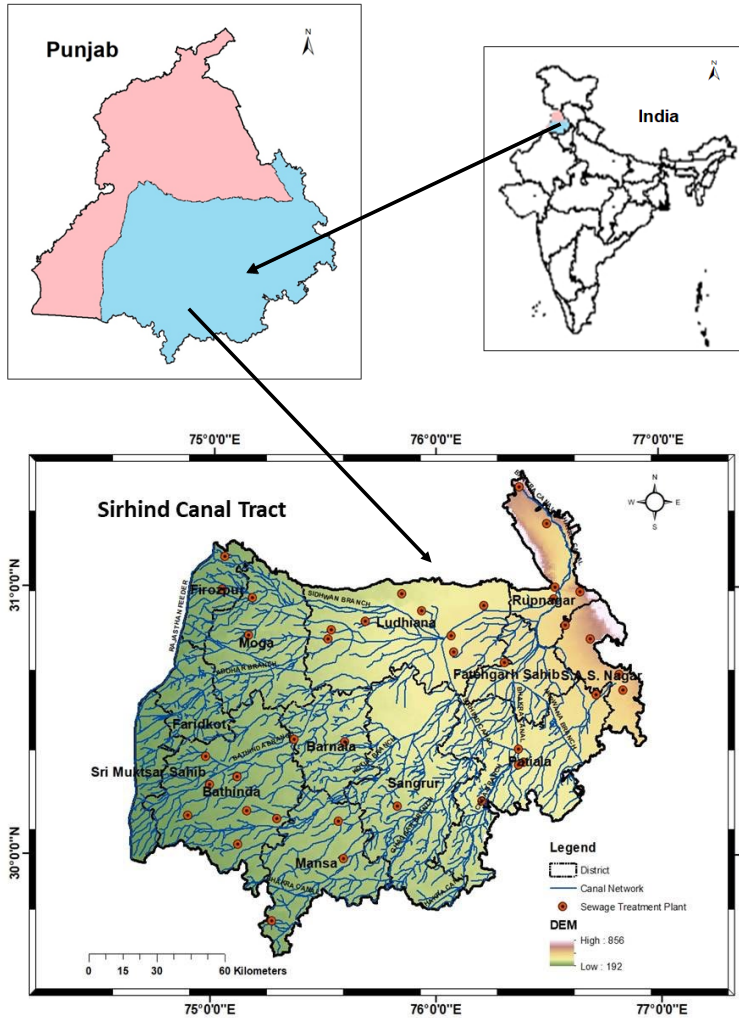
191 **Figure 1.** Simulation modelling workflow for the evaluation of alternative groundwater
192 management strategies.

193

194 **3.1 Case Identification**

195 The Sirhind Canal Tract, known as Cis-Doab, is one of the most productive aquifer
196 systems in the Indo-Gangetic basin. The tract lies to the south of the perennial Sutlej river
197 between latitude 29° 53 ´ N and 31° 37 ´ N, and longitude 74° 50 ´ E and 76° 51 ´ E. The 2.60
198 million ha area of the tract covers the districts of Barnala, Bathinda, Fatehgarh Sahib,
199 Ludhiana, Mansa, Mohali, Moga, Patiala, Ropar, and Sangrur with some area of Ferozpur,
200 Faridkot and Muktsar (Figure 2) entirely. The elevation of the tract varies from 340 m in the
201 northeast to 190 m above the mean sea level in the southwest (Marok et al. 2000). The land
202 slopes from northeast to southwest with an average gradient of 0.68 m per 1,000 m. Alluvium
203 deposits of the Sutlej River, the Ghaggar River and their tributaries form the principal soil in
204 the tract.

205



206

207 **Figure 2.** The Sirhind Canal Tract, Punjab State, India [based on data retrieved from the
 208 Punjab Remote Sensing Centre, Ludhiana].

209

210 **3.2 Geographic Area and Climate Determination**

211 There are two main cropping seasons in the tract, namely *Kharif* (summer) and *Rabi*
 212 (winter), with rice being the main cultivated crop during the *Kharif* season and wheat the
 213 leading crop during the *Rabi* season. Wide seasonal variations in temperature characterise
 214 the climate in the region. In particular, summer lasts from April till October with the

215 temperature ranging between 21-51°C, whereas the winter season extends from October till
216 April with the temperature ranging between 7-27°C. The mean air temperature in the tract
217 during winter is about 5°C, while during summer, it reaches 40°C, monthly. In addition, the
218 mean monthly relative humidity reaches 90% during monsoon (i.e., July to September) and
219 varies between 30-60% during summer (Marok et al. 2000). Furthermore, the eastern side of
220 the tract receives a normal annual rainfall of 1,000 mm or more, whereas it declines down
221 to less than 400 mm towards the west (Miglani et al. 2015). Overall, significant variations in
222 temperature, relative humidity and rainfall can be observed across the tract.

223 The tract comprises an extended canal network, from main canals to distributaries
224 and other inland waterways. The Bhakra Main Line Canal and the Sirhind Canal are the two
225 primary water sources for irrigation in the area. The Sirhind Canal further diverges to the
226 Sidhwan branch, the Abohar branch, the Bathinda branch and **the First and Patiala feeders**.
227 The First Patiala feeder further bifurcates into the Kotla branch and the Second Patiala
228 feeder. The latter feeder is subdivided into the Ghaggar branch and the Third Patiala feeder
229 (Figure 2). The inter-state canal system of Bhakra Main Line is used to irrigate the eastern
230 and southeastern part of Punjab, along with the northern parts of Haryana and Rajasthan,
231 through the Nirwana branch. The Sirhind Canal system is used to irrigate the central and
232 western parts of the tract (Marok et al. 2000).

233

234 **3.3 Groundwater Simulation**

235 The groundwater simulation was developed via employing the MODFLOW model
236 (McDonald and Harbaugh, 1988), in a Processing MODFLOW for Windows environment. The

237 MODFLOW model utilises a partial differential equation to quantify the three-dimensional
238 movement of groundwater in a heterogeneous and anisotropic aquifer (Equation 1):

$$239 \quad \frac{\partial}{\partial x} \left(K_{xx} \frac{\partial h}{\partial x} \right) + \frac{\partial}{\partial y} \left(K_{yy} \frac{\partial h}{\partial y} \right) + \frac{\partial}{\partial z} \left(K_{zz} \frac{\partial h}{\partial z} \right) \pm w = S_s \frac{\partial h}{\partial t} \quad \text{Eq. (1)}$$

240 where, K_{xx} , K_{yy} and K_{zz} represent the hydraulic conductivity along the x , y and z
241 coordinate axes in m/day; h is the hydraulic head in m; w is the volumetric flux per unit
242 volume and represents sources and/or sinks of water in m³/day; S_s denotes the specific
243 storage of the porous material per m; and t is time in days.

244

245 3.3.1 Conceptualisation

246 To create a conceptual model of the tract, data on the lithology of the aquifer (i.e.,
247 structural contours, bore logs, aquifer properties), piezometer levels, climatic data, aquifer
248 abstractions, canal data, canal network, and their L-sections were collected. Principal
249 aquifers of sand beds, separated by clay beds at various depths, characterise the
250 lithological heterogeneity in the area. Unconsolidated alluvium of the tract was accounted
251 with beds of gravel and cemented sand in multiple locations (Marok et al. 2000). Phreatic
252 and confined aquifer conditions in the alluvial material comprised of sand, silt, clay and
253 kankar formulate the medium for groundwater storage. The aquifer properties such as
254 hydraulic conductivity and specific yield were estimated indirectly from the logs of 500 wells
255 scattered across the tract (Figure3) using the following analytic descriptions provided by
256 Todd (1980) via Equations 2 and 3:

$$257 \quad K_h = \frac{\sum K_{hi} \times d_i}{\sum d_i} \quad \text{Eq. (2)}$$

258
$$S_y = \frac{\sum S_{yi} \times d_i}{\sum d_i} \quad \text{Eq. (3)}$$

259 where,

260 K_h = hydraulic conductivity,

261 S_y = specific yield,

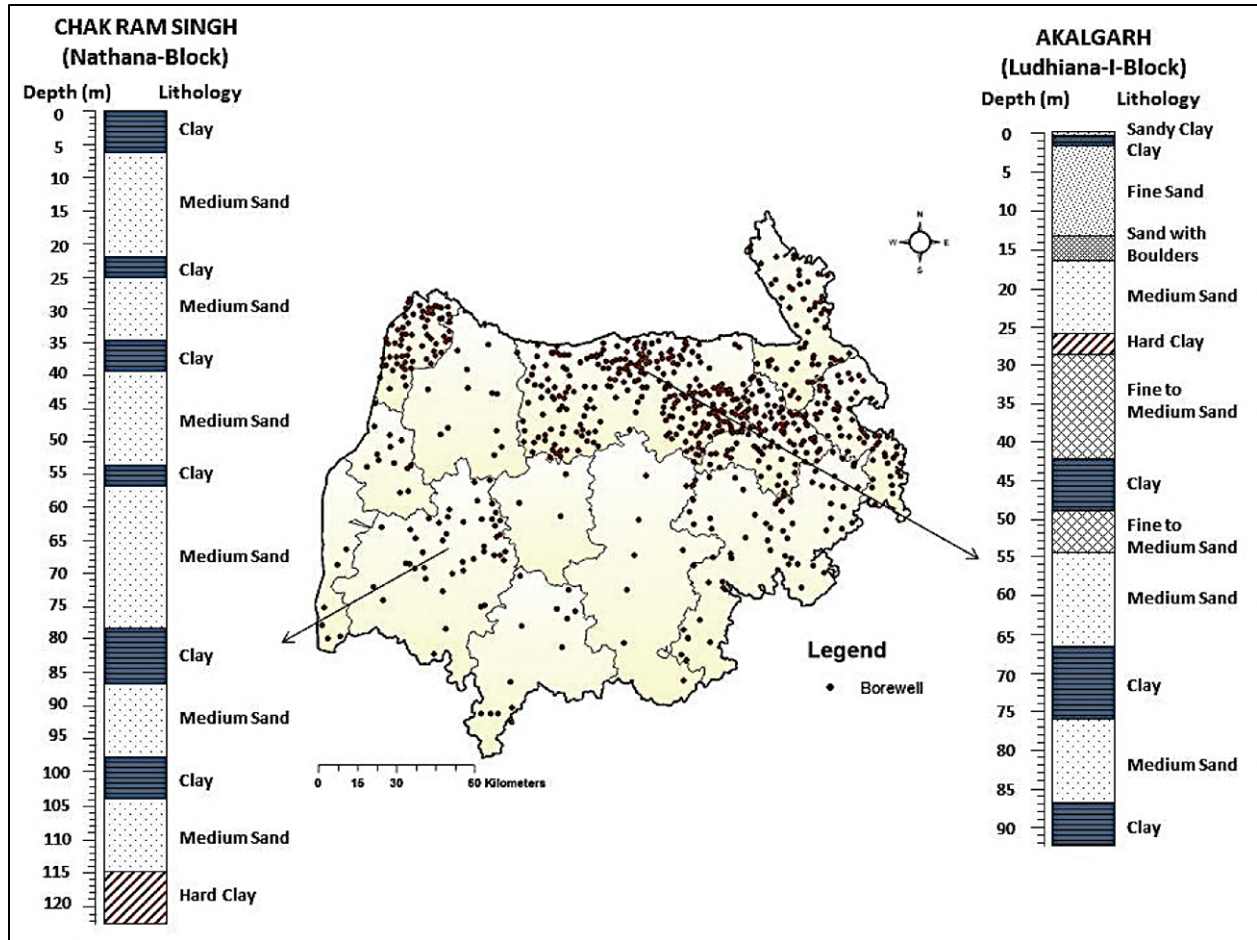
262 K_{hi} = hydraulic conductivity of the layer,

263 S_{yi} = specific yield of the layer,

264 d_i = depth of the layer.

265 The calculated hydraulic conductivity and specific yield were subsequently
266 optimised using the PEST automated nonlinear parameter estimator (Doherty, 2016).
267 Primary time series data describing the water level of more than 200 observation points for
268 the Sirhind Canal Tract for the pre-monsoon period (i.e., June) was collected from the
269 Department of Agriculture and Water Resources and Environment Directorate, Punjab. The
270 data was used to calculate the hydraulic heads of the tract to calibrate and validate the
271 model.

272



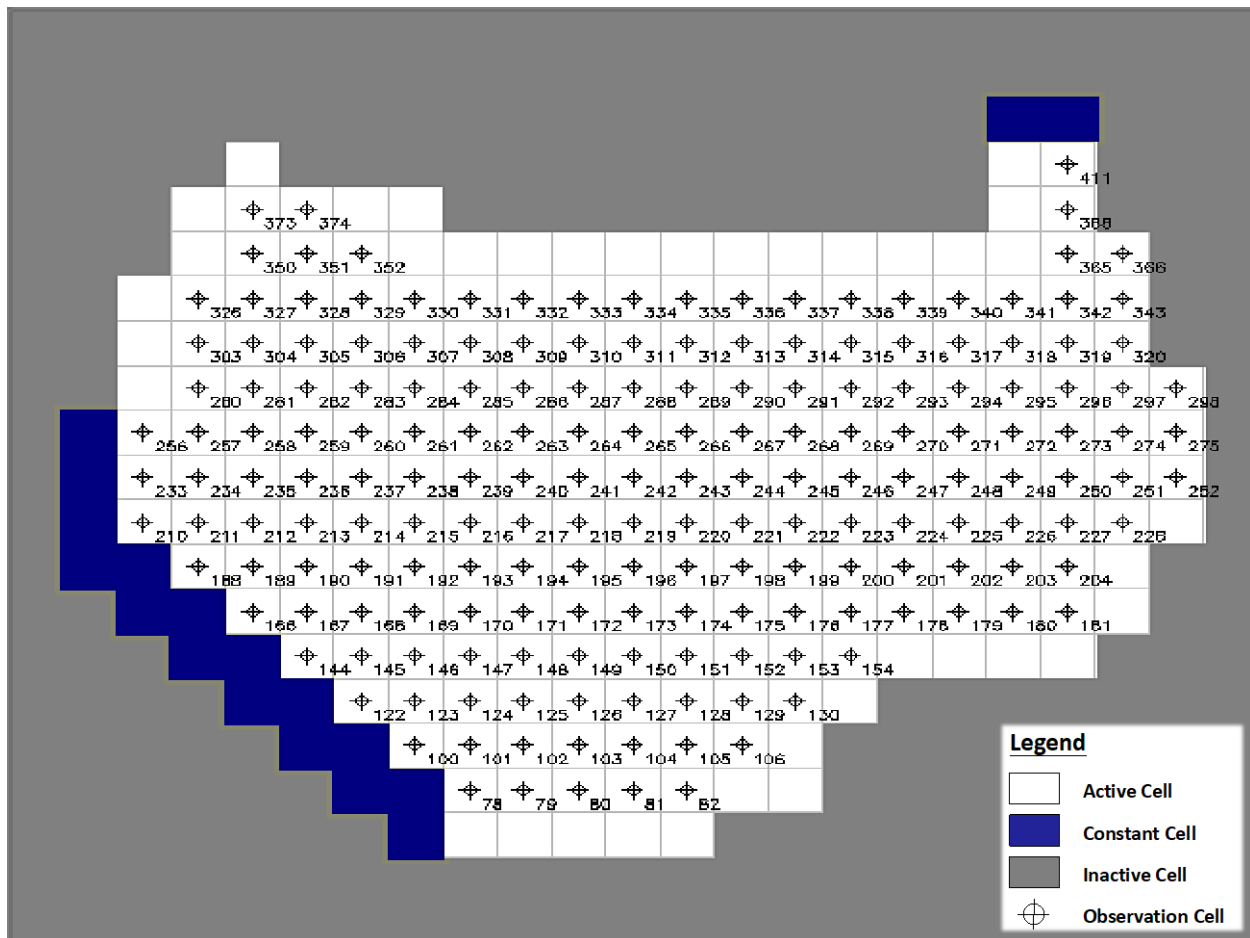
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274 **Figure 3.** Location and lithological description of observation points [based on data
 275 retrieved from the Punjab Remote Sensing Centre, Ludhiana].

276

277 **3.3.2 Discretisation**

278 Regionally, the aquifer behaves like a water-table aquifer and extends approximately
 279 220 km in length and 220 km in breadth (north-south orientation). After the conceptual
 280 model was developed, the area was discretised into 21 rows and 23 columns with a constant
 281 grid spacing of 10×10 km, considering an unconfined aquifer up to a depth of 300 m,
 282 resulting in a total of 483 discretised cells (Figure 4). The digital map of the area was
 283 imported in MODFLOW to represent the real-world location of the study area.



285

286 **Figure 4.** Discretisation of the Sirhind Canal Tract using MODFLOW.

287

288 **3.3.3 Boundary conditions demarcation**

289 Hydrologically, the Shivalik foothills delimit the tract towards east and north-east,
 290 the Ghaggar River flows in the southern boundary, the Bhakra Main Line canal courses
 291 towards the south-west and the Rajasthan feeder streams in the west. However, based on
 292 historical annual groundwater level fluctuations, the cells lying southwest of the tract were
 293 considered constant head cells, i.e., the storage term was not considered, and the head
 294 specified in the beginning was kept constant throughout the simulation. All other cells lying

295 inside the region were simulated as active cells for which hydraulic heads were computed
296 throughout all time steps of the simulation. The flux from/to individual boundary cells was
297 computed using Darcy's law. The cells lying outside the boundaries of the study area were
298 considered inactive or no flow cells meaning that no inflow/outflow of the cell was permitted
299 in any time step of the simulation. Out of the total 483 discretised cells for the model, 226
300 were marked as active cells, 18 as constant head cells, and 236 as dormant cells. The study
301 area was simulated using 188 cells (observation cells) out of 226 active cells. The model was
302 run in a transient state to simulate the annual changes in groundwater levels. A stress period
303 of 365 days was selected to simulate the annual effects of recharge and draft on the
304 groundwater system. The temporal parameters, i.e., recharge and draft, were simulated
305 using well and recharge flow packages for each discretised cell.

306

307 *3.3.4 Groundwater draft and recharge estimation*

308 The number of tubewells in each block (smallest administrative unit) was multiplied
309 with their values of unit draft to calculate the groundwater pumped during the period from
310 1998-1999 to 2017-2018 (Ministry of Water Resources, 2017). However, the computed draft
311 was increased or decreased by 10% while considering the variability of rainfall as $\pm 19\%$ or
312 more compared to average rainfall during wet/dry years. The total groundwater recharge for
313 the tract, during 1998-2018, was estimated using Equation 4:

$$314 \quad R_g = R_r + R_c + R_{cia} + R_{tia} - E_t \quad \text{Eq. (4)}$$

315 where,

316 R_g = total groundwater recharge (mm),

317 R_r = recharge to groundwater from percolation of rainfall (mm),
 318 R_c = recharge to groundwater through seepage from the main canals,
 319 branches and their distributaries (mm),
 320 R_{cia} = recharge to groundwater due to return flow from canal irrigated area
 321 (mm),
 322 R_{tia} = recharge due to return flow from tubewell irrigated areas (mm),
 323 E_t = evaporation from shallow water table areas (capillary rise from shallow
 324 groundwater into the unsaturated zone and contribution to evaporation)
 325 (mm),

326 and the net recharge was evaluated via using Equation 5:

$$327 \quad R_t = R_g - Q_p \pm Q_g \quad \text{Eq. (5)}$$

328 where,

329 Q_g = groundwater inflow/outflow to/from the area from/to the neighbouring
 330 areas (mm),

331 Q_p = groundwater abstraction (mm).

332 Due to the negligibility of shallow water tables in the study area, the evaporation was
 333 considered to be insignificant. As per recommended norms by the Ministry of Water
 334 Resources (2017), the recharge from rainfall (R_r) was considered 25% of the total rainfall.
 335 Also, no recharge was assumed in case monthly rainfall collected in the region was
 336 accounted for less than 50 mm (Ministry of Water Resources, 2017). The seepage from the
 337 canal network (R_c) for the unlined canal was taken as 18 ha-m/day/106 m². Seepage losses

338 were assumed to be 20% of the aforementioned value for the lined canals. Although an
339 extended webbed canal system characterises the region, only 263 mm (around 27%) of
340 irrigation water is applied through it. A value of 20% of water delivered at the outlet for
341 application in the field was considered as recharge from the canal irrigated areas for the
342 non-paddy area. In contrast, 35% of the water delivered at the outlet was considered
343 recharge for the paddy area. Similarly, the return flow from tubewell irrigated areas was
344 considered to be 25% from non-paddy fields and 35% from paddy fields.

345

346 **3.4 Model Calibration and Validation**

347 Long term data is fundamental for the accurate calibration of a groundwater model
348 (Kaur et al. 2015). In particular, the model in this research was calibrated with data for the
349 period 1998-2013; the hydraulic heads predicted by the model were matched with the
350 historic field-measured values by adjusting aquifer parameters within plausible ranges.
351 Thereafter, the calibrated flow model was validated for the period 2013-2018. The degree of
352 fit between model simulations and field measurements was quantified by using the Mean
353 Error (ME) and the Root Mean Square Error (RMSE) given by Equation 6 and Equation 7,
354 respectively:

$$355 \quad ME = \frac{1}{n} \sum_{i=1}^n (h_o - h_s) \quad \text{Eq. (6)}$$

$$356 \quad RMSE = \sqrt{\frac{1}{n} \sum_{i=1}^n (h_o - h_s)^2} \quad \text{Eq. (7)}$$

357 where,

358 h_o = observed hydraulic head,

359 h_s = simulated hydraulic head,

360 n = number of observed years.

361

362 **3.5 Simulation of Groundwater Management Strategies, Energy Requirements and** 363 **Carbon Emissions**

364 The validated model was then used to predict future groundwater levels by 2025 and
365 2030. Specifically, four different pumping scenarios were examined in combination with
366 three irrigation conditions (Table 2), resulting in a total of twelve alternative groundwater
367 management strategies (i.e., “Irrigation Condition – Pumping Scenario” combinations).

368

369 **Table 2.** Groundwater management strategies.

370 [Table 2 about here]

371

372 There are 42 STP plants geographically scattered across the Sirhind Canal Tract
373 (Figure 2) which can be used to provide treated wastewater for irrigation (Punjab Pollution
374 Control Board,2017). The wastewater treatment capacity of these operating STPs, which
375 can be exploited for irrigation usage, is tabulated in Table A1 in Appendix I. Acceleration of
376 industrialisation and urbanisation has increased the urban sewage discharge rates at an
377 annual rate of 5%. However, the current capacity for sewage treatment is relatively low (Yang
378 et al. 2017). Notably, the total STPs’ wastewater treatment capacity in the study area is
379 considerably less compared to the potential wastewater generated from the 66 blocks of
380 the area (Table A2 in Appendix I). Although certain progress has been made regarding the

381 exploitation of existing sewage treatment facilities, their geographical dispersion is a
382 challenge (Cheng et al. 2020). To that end, new STPs need to be constructed to address the
383 volumetric gap in wastewater treatment capacity.

384 While computing groundwater levels under irrigation conditions B and C
385 incorporated wastewater augmentation, groundwater abstraction was reduced from the
386 corresponding cells receiving additional irrigation water. Management of water resources
387 was anticipated by comparing the predicted groundwater levels of the twelve strategies
388 formulated for this research. The groundwater levels projected under the different irrigation
389 conditions and pumping scenarios were finally analysed to determine comprehensive
390 policy-making recommendations for the environmentally sustainable management of the
391 groundwater table in the study area by the year 2030.

392 Rural electrification reflects upon the fact that the energised tubewells have
393 increased from 0.19 million in 1970 to 1.47 million in 2017 (ESOPB 2020, p.162). Therefore,
394 the depletion of groundwater level directly and indirectly impacts energy, cost, and carbon
395 emissions. Kaur et al. (2016) discussed that a decline in groundwater level by 5.47 m (1998-
396 2012) increased energy requirements by 67% and carbon emissions by 110%. In addition,
397 Dhillon et al. (2018) estimated an energy requirement of 7,919.60 mega kilowatt-hours
398 (MkW-h) for groundwater pumping in central Punjab, resulting in 1,349.6 thousand tons of
399 carbon emissions. In the same vein, the energy requirements and CO₂ emissions were
400 estimated for the tract under the investigated management strategies.

401

402 **4. Results and Discussion**

403 The groundwater simulation model generated interesting results. The following
404 subsections present the modelling calibration and validation results while discussing the
405 analysis output per investigated management strategy.

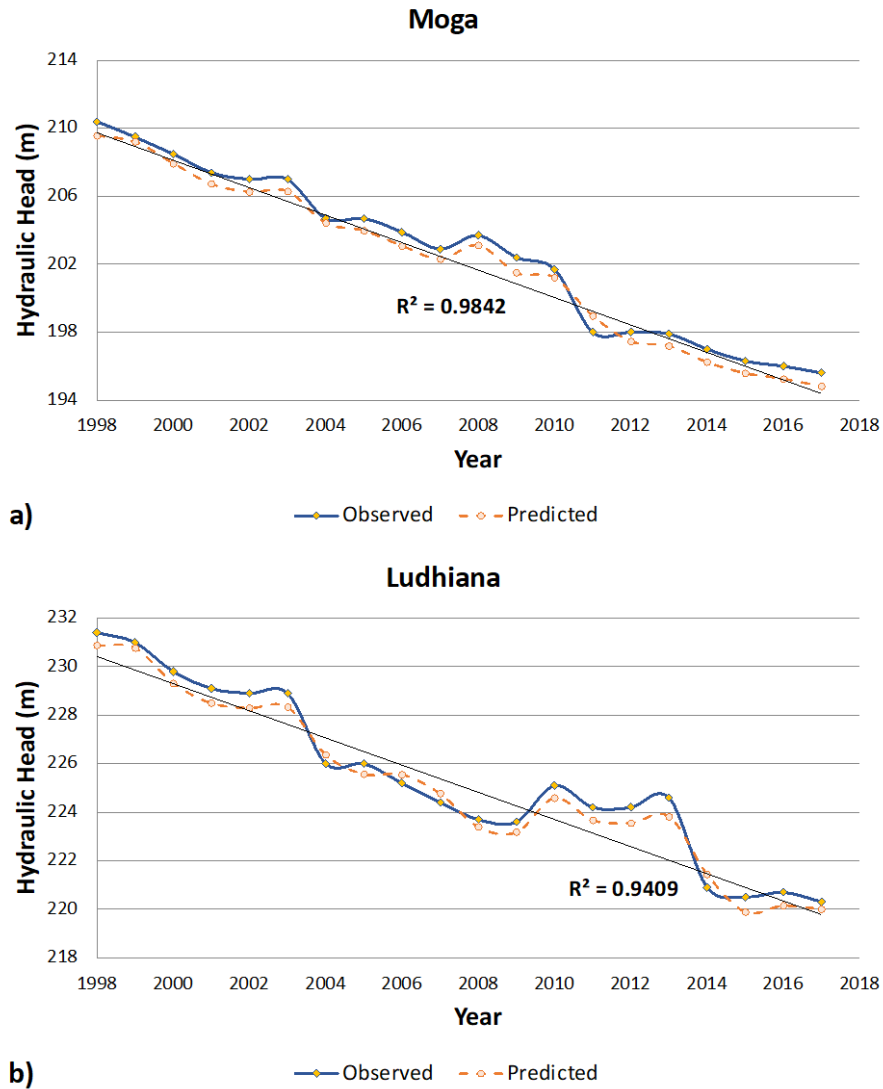
406

407 **4.1 Calibration and Validation**

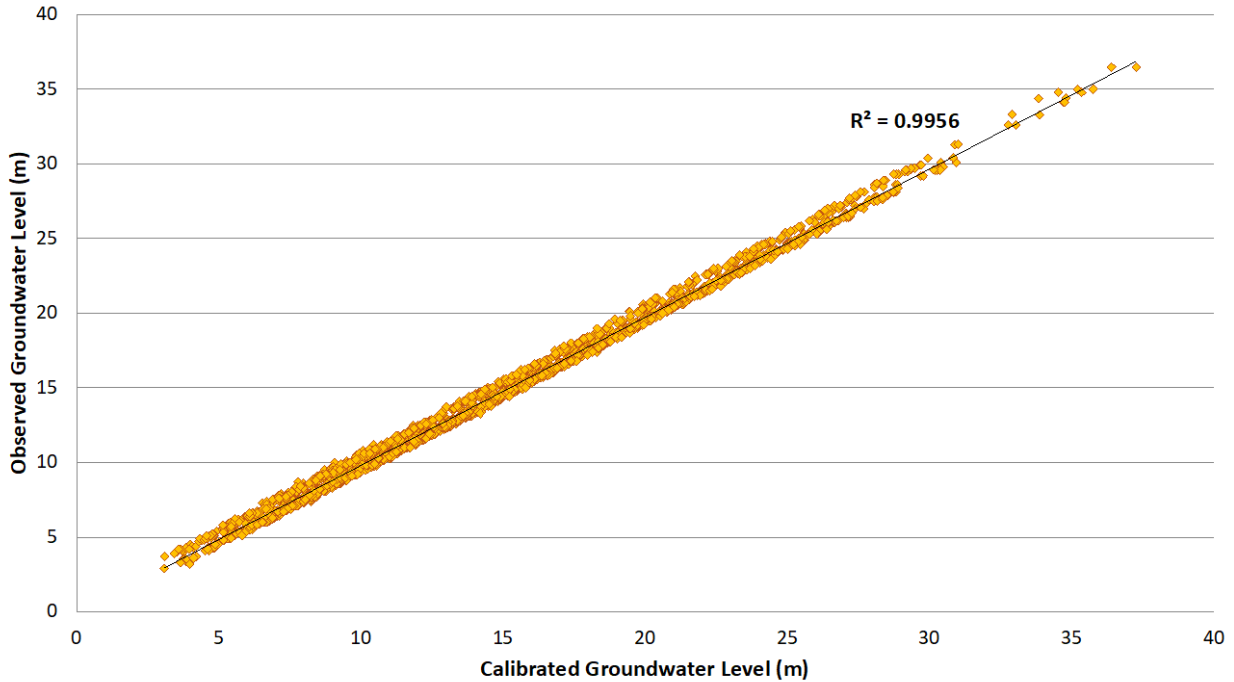
408 The groundwater model reasonably simulated the significant spatial and temporal
409 variation (due to changes in recharge and draft) in the magnitude of hydraulic heads (Figure
410 5). The results of the calibrated model, presented in Figure 6, demonstrate a good fit
411 between the simulated hydraulic heads and the real-world observations. Due to the
412 heterogeneous topography and lithology of the region, the calibrated hydraulic conductivity
413 was found to range from 1 m/day to 100 m/day, whereas the specific yield varied from 0.02
414 to 0.2. The average difference between the observed and the simulated water levels for most
415 internal nodes varied between 0 and 0.75 m. Furthermore, the calculated average ME was
416 0.227 during calibration and 0.341 during validation, while the average RMSE was calculated
417 0.470 during calibration and 0.662 during the validation period. All the calibration and
418 validation results for the Sirhind Canal Tract are considered reasonable and within an
419 acceptable range. For deviations within these limits, the performance of a groundwater
420 model can be regarded as satisfactory (Toews and Allen, 2009).

421 The calculated groundwater abstraction in the tract for 2018 was 27.30 billion m³,
422 and the total recharge was 13.80 billion m³. Notably, groundwater abstraction in 2018 was
423 higher by 77% compared to 1998 due to an increase in the number of tubewells. Net

424 recharge function of rainfall, irrigation and groundwater storage and abstraction ranged
425 from -60.00 to +32.40 cm in different years with an average of -10.30 cm.
426



427
428 **Figure 5.** Comparison between observed and predicted hydraulic heads at the Districts of
429 Moga and Ludhiana at the Sirhind Canal Tract, as part of the model calibration process.
430



431

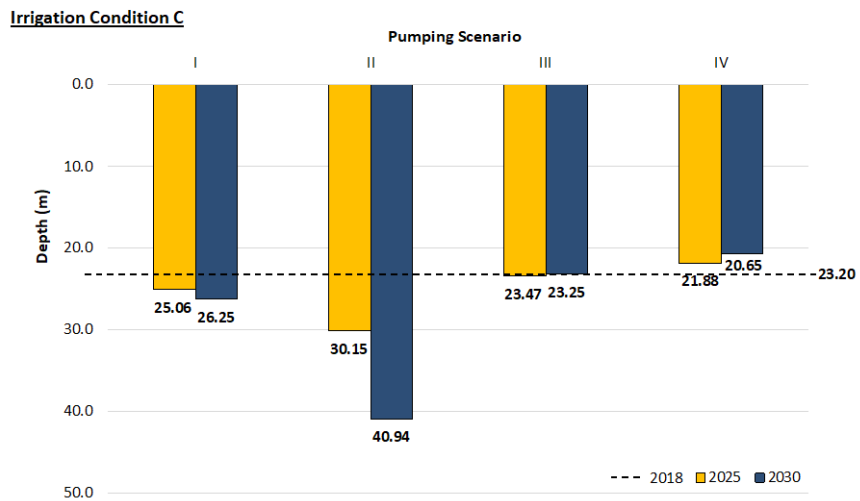
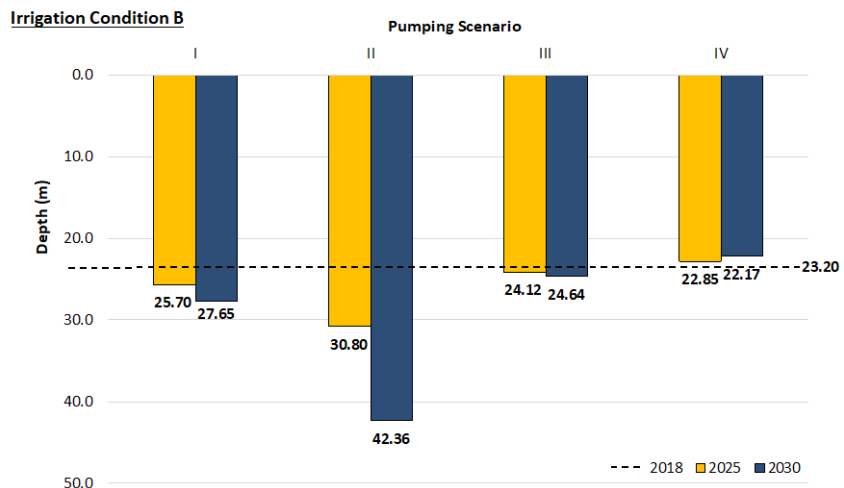
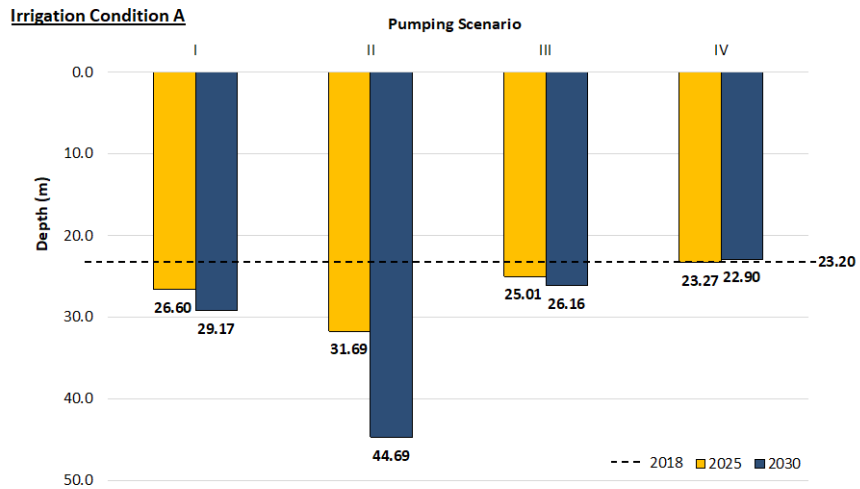
432 **Figure 6.** Measured and calibrated groundwater levels at the Sirhind Canal Tract.

433

434 **4.2 Simulation**

435 The water table at the Sirhind Canal Tract, observed to be 23.20 m in 2018, was
 436 simulated for the years 2025 and 2030 under three irrigation conditions, namely: (i) normal
 437 irrigation without any augmentation; (ii) augmentation of irrigation with treated wastewater
 438 from existing STPs; and (iii) augmentation of irrigation with potentially treated wastewater.
 439 The research findings in terms of projected groundwater level under the considered
 440 pumping scenarios and irrigation conditions are summarised in Figure 7 and exemplified
 441 below.

442



443

444 **Figure 7.** Projected groundwater levels under alternative pumping scenarios and irrigation

445 conditions.

446

447 *4.2.1 Irrigation Condition A: Normal irrigation without any augmentation*

448 In the normal irrigation condition, groundwater levels are expected to fall by 3.40 m
449 and 5.96 m in 2025 and 2030, respectively, under Scenario I. Furthermore, because tubewell
450 density has been increasing in the region, the future groundwater abstraction of 33.2 billion
451 m³ in 2025 and 34.9 billion m³ in 2030 under Scenario II will increase to 32.60 billion m³ and
452 34.20 billion m³ by the years 2025 and 2030, respectively, causing the water table to decline
453 by 8.49 m in 2025 and 21.49 m in 2030. In a similar scenario by Shakoor et al. (2018),
454 depletion of 14.00 m in groundwater level was reported for the District of Bhawana in Punjab,
455 Pakistan. In Scenario III, with increased canal water supplies, the predicted fall in
456 groundwater level would be 2.95 m by 2030. In Scenario IV, where groundwater abstraction
457 is proportionally reduced, demand for irrigation water is being compensated by additional
458 canal water supplies; thus, it is predicted to make groundwater table rise by 0.30 m in 2030.

459

460 *4.2.2 Irrigation Condition B: Augmentation of irrigation with treated wastewater from existing*
461 *Sewage Treatment Plants*

462 In irrigation Condition B, the groundwater level under Scenario I is projected to fall by
463 2.50 m in 2025 and 4.44 m in 2030. The fall in the water table estimated for the management
464 strategy “Condition B – Scenario I” is 3.37% and 5.19% less than the water table expected
465 to fall by 2025 and 2030, respectively, under “Condition A – Scenario I”. Although the
466 tubewell density is considered to increase in Scenario II, the augmentation with treated
467 wastewater from STPs would decrease the groundwater abstraction approximately by

468 5.17%, and the water table is expected to fall by 2.82% in 2025 and 5.22% in 2030 compared
469 to the water levels estimated in strategy “Condition A – Scenario II”. In Scenario III of surplus
470 canal water supply along with no additional tubewell drilling, the water table would
471 experience a fall of 0.92 m in 2020, resulting in an improvement of groundwater levels until
472 2030. Under pumping Scenario IV, reduced groundwater abstraction, additional canal water
473 supply and wastewater augmentation from existing STPs would produce a rise in the
474 groundwater levels by 1.51% in 2025 and 4.43% in 2030 compared to the reference value of
475 23.20 m observed in the year 2018. Moreover, the estimated levels are 1.80% and 3.18%
476 higher than the levels estimated under strategy “Condition A – Scenario IV” by 2025 and
477 2030, respectively.

478

479 *4.2.3 Irrigation Condition C: Augmentation of irrigation with potentially treated wastewater*

480 The results estimated under irrigation Condition C could be regarded as the
481 promising groundwater levels produced by exploiting wastewater along with diverse
482 pumpage and canal water utilisation. Notably, in Condition C, the water table level would
483 fall by 1.85 m in 2025 and 3.05 m in 2030 under the pumping Scenario I. These levels
484 obtained in the management strategy “Condition C – Scenario I” are 5.80% and 9.99% higher
485 than the levels estimated for the years 2025 and 2030, respectively, during the strategy
486 “Condition A – Scenario I”. Under Scenario II, groundwater abstraction of 30.72 billion m³ in
487 2025 and 32.3 billion m³ in 2030 would cause the water table to rise by 4.87% and 8.39% by
488 2025 and 2030, respectively, compared to the predicted values under management strategy
489 “Condition A – Scenario II”. Furthermore, in Scenario III, the water table after possible

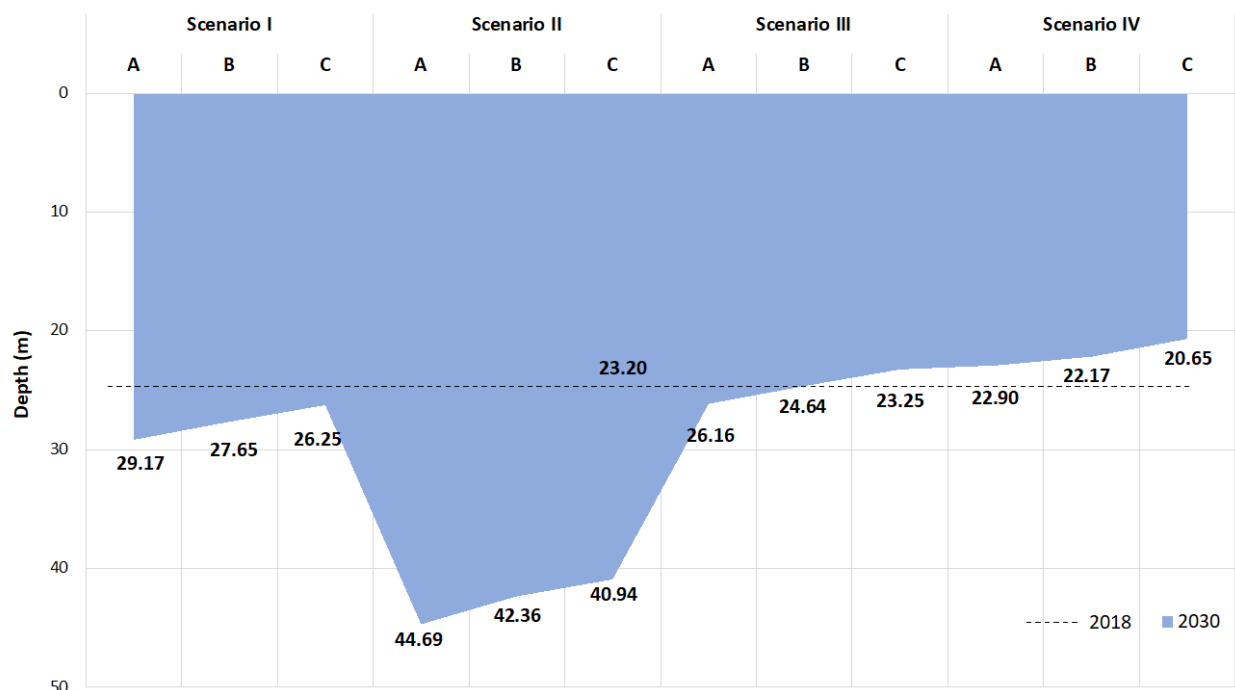
490 augmentation of wastewater would fall by only 1.16% in 2025 and 0.29% in 2030 from the
491 reference water table in 2018, which is 6.14% and 11.12% higher than the groundwater
492 levels predicted for the years 2025 and 2030, respectively, under the management strategy
493 “Condition A – Scenario III”. In pumping Scenario IV, potential wastewater augmentation
494 along with increased canal water supplies and decreased groundwater abstraction could
495 cause the water table to rise by 5.67% and 10.97% by 2025 and 2030, respectively,
496 compared to the water table depth of 23.20 m calculated for the year 2018. These results
497 are 5.95% and 9.80% higher than predicted levels for 2025 and 2030, respectively, under the
498 management strategy “Condition A – Scenario IV”.

499

500 **4.3 Groundwater Levels under Alternative Management Strategies**

501 A comparison across pumping scenarios and irrigation conditions indicates that it is
502 feasible to completely reverse the falling groundwater trend by 2030 in Scenario IV (Figure
503 8). Indeed, incorporation of all the strategies (i.e., enhanced canal water supplies,
504 decreased groundwater abstraction and potential wastewater augmentation with irrigation
505 under condition “Condition C – Scenario IV”) could uplift the groundwater level at 20.7 m in
506 2030, which is 10.97% higher than the level at 23.20 m predicted for the year 2018. In
507 Scenario II, which can be considered the worst-case scenario, none of the proposed
508 management strategies will be effective for environmentally sustainable groundwater
509 management. Augmentation of irrigation supplies with treated wastewater from existing
510 STPs (i.e., Condition B) seems to be a realistic option and could ensure an improvement in
511 the groundwater levels in the year 2030 by 5.19%, 5.22%, 5.80% and 3.18% under pumping

512 Scenarios I, II, III and IV, of the management strategy “Condition A”. Additionally, in case all
 513 wastewater generated in future is treated (i.e., Condition C), groundwater levels are
 514 expected to improve by 9.99%, 8.39%, 11.12% and 9.80% in 2030 under pumping Scenarios
 515 I, II, III and IV, compared to the management strategy “Condition A” (Figure 8).
 516



517
 518 **Figure 8.** Comparison of expected groundwater levels in 2030 under the investigated
 519 management strategies.
 520

521 **4.4 Energy Requirements and Carbon Emissions**

522 Groundwater pumping in Punjab interacts with energy consumption, thus dictating
 523 the need to include energy in relevant studies to promote integrated resources management
 524 and policy coherence (Wichelns, 2017). In this regard, the energy requirements and CO₂

525 emissions estimated for the tract under the different investigated management strategies
526 are summarised in Table 3.

527

528 **Table 3.** Energy requirements and CO₂ emissions under the investigated groundwater
529 management strategies.

530

531 [Table 3 about here]

532

533 In this research, it was estimated that 5,945.1 Mkw-h of energy is required to pump
534 27.3 billion m³ of groundwater which resulted in 1,507.9 thousand tons of CO₂ emissions in
535 the year 2018 for the Sirhind Canal Tract. During the simulation of the alternative
536 management strategies, it was found that the energy requirement vis-a-vis CO₂ emissions
537 would be the highest (i.e., 2,825.7 thousand tons) under the strategy “Condition A – Scenario
538 II”. The carbon emissions estimated for the tract during the other strategies would vary by
539 11-17% from the highest emissions strategy.

540

541 **5. Conclusions**

542 The increased reliance of farmers in the Indian Punjab on groundwater supplies for
543 irrigation purposes has been the major cause of groundwater decline in the Sirhind Canal
544 Tract. Nearly 99% of the State’s net agricultural area is sown under irrigation, while surface
545 water resources accommodate only 28% of the total irrigation needs and the balance
546 abstracted through tubewells (ESOPB 2020, p.157). This research highlights the importance

547 of groundwater models for predicting groundwater levels under alternative management
548 scenarios. Overall, the findings of this research indicate that new tubewells for farm
549 irrigation purposes should be avoided and surface water exploitation promoted.

550

551 **5.1 Academic Contributions**

552 This study contributes to the extant body of literature by investigating alternative
553 irrigation, distribution and abstraction strategies through validated simulation models using
554 both primary and secondary datasets. This is one of the first studies in the Punjab region that
555 confirms the severity of current and projected groundwater levels with quantified analysis
556 of the Sirhind Canal Tract in terms of potential mitigation through alternative irrigation,
557 distribution, and abstraction strategies. Specifically, the Sirhind Canal Tract MODFLOW
558 simulation model suggests that canal water supplies could be beneficial to reverse
559 groundwater level decline and help increase the water level by 11% above that in the year
560 2018. Furthermore, the usage of treated wastewater in irrigation and the different
561 groundwater pumping scenarios provides promising scenarios that reverse recent declines
562 in groundwater water resources in the tract. In addition, in terms of energy consumption
563 involved in tubewell abstraction, the suggested alternative approaches would reduce water
564 abstraction leading to reduced CO₂ emissions of approximately a million ton by 2030. The
565 originality of this research lies in the integration of primary and secondary data over an
566 extensive geography highly exposed to overexploitation, the strength of model validation,
567 the generation of quantified alternative groundwater management strategies and the
568 consequent Environmental, Social and Governance (ESG) impact.

569

570 **5.2 Policy-making Implications**

571 The simulation modelling investigation successfully captured the spatial and
572 temporal groundwater behaviour, and the analysis of the results focused on the decrease in
573 groundwater abstractions. Specifically, the developed MODFLOW model was applied to
574 predict future groundwater levels by 2030 under twelve management strategies. The
575 research findings can help articulate recommendations to inform pertinent policy-making
576 decisions based on investigating the impact of alternative instruments, as proposed in the
577 literature (Moors et al. 2011).

578 Firstly, as per the historical trend of groundwater abstraction, the model's findings
579 suggest a steep decline of 21.49 m in groundwater level by 2030. During any examined
580 irrigation conditions, an increase in the number of tubewell scan results in a higher water
581 table fall. Even in case the existing tubewells are allowed to continue operating with
582 prohibitions on developing new tubewells, the State will be facing a decline of 5.96 m in
583 2030, which is still not considered a sustainable pumping scenario.

584 Secondly, the findings further suggest that an increase of 10% in canal water supplies
585 can prove beneficial to reverse groundwater level decline. The canal water available in
586 Punjab can be easily exploited to irrigate 3.08 million ha, out of the total 4.29 million ha of
587 the State's land area. **This is rational in a way that the canal irrigated area has been reduced
588 by 10% (see Appendix II).**

589 Thirdly, in alignment with policy findings from Vijet al. (2021), this research argues
590 that the many STPs which exist in the vicinity of cities/towns for providing environmentally

591 safe sewage water can augment substantial portions of irrigation water supplies.
592 Substantiating the irrigation water supplies with treated wastewater from existing STPs
593 reduces the surplus groundwater abstraction by 3-4%, which can be further enhanced up to
594 7-11% by potentially treating the entire volume of wastewater generated in the region. This
595 would prove beneficial to restore the groundwater level by 10.97%. Moreover, deficit
596 groundwater abstraction and surplus wastewater augmentation and canal water supplies
597 under "Condition C-Scenario IV" will reduce energy demand and lead to reduced CO₂
598 emissions of approximately 966.6 thousand tons by 2030.

599

600 **5.3 Limitations and Future Research**

601 This modelling research should be considered as a first step to quantify the
602 environmental sustainability of groundwater resources via leveraging surface water and
603 reuse options. In the study, it is assumed that canal water supplies may increase in the
604 future; however, the decline in area under canal irrigation could decrease the water flow in
605 the rivers. This needs to be scientifically assessed if there is a decrease in the river water
606 flow and can be crucial material evidence for Punjab's case in the river water dispute
607 pending in the Supreme Court (Singh 2018) and accordingly, share allocations to
608 neighbouring states may be proportionately reduced. The analysis focuses exclusively on
609 groundwater quantity issues without considering the groundwater quality component due
610 to a lack of timeseries data. Also, certain aspects like reduced crop water demand through
611 crop diversification, improved field water use efficiency as in direct-seeded rice and micro-
612 irrigation techniques, and managed aquifer recharge programmes, which significantly

613 impact groundwater allocation for irrigation and recharge, were not considered in this study.
614 Including such parameters provides research avenues for future studies to guide an
615 integrated groundwater policy-making agenda in the region. Incorporation of groundwater
616 modelling to end-to-end food supply networks that consider the water footprint of products
617 could further provide a more comprehensive view of the sustainability impact of both
618 upstream and downstream operations (Aivazidou et al., 2018). Such a comprehensive
619 systems view could further inform governance decisions for ensuring equitable groundwater
620 extraction in support of marginalised and small farmers (Hoogesteger and Wester, 2015).

621 The number of installed tubewells provided a reasonable estimation of groundwater
622 use for agricultural purposes. The model was able to effectively capture the annual trends
623 of groundwater levels, which can be utilised for a broader policy formulation. However, this
624 method does not adequately address domestic and industrial groundwater usage. Improved
625 data availability (through mass monitoring and telemetry systems) would allow modelling at
626 a more granular spatial and temporal resolution to give insights into the groundwater flow
627 and recharge processes.

628

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630 [The Acknowledgements are included in the 'Title Page' file, uploaded separately, for
631 ensuring anonymity during the review process].

632

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816

817 **Appendix I**

818

819 **Table A1.** Wastewater treatment capacity of Sewage Treatment Plants located across the
820 Sirhind Canal Tract [Source: Punjab Pollution Control Board (2017)].

821 [Table A1 about here]

822

823 **Table A2.** Wastewater generation in the Sirhind Canal Tract [Source: Punjab Pollution
824 Control Board (2017)].

825 [Table A2 about here]

826

827 **Appendix II**

828 **Table A3.** Canal irrigated area in Punjab.

Area	Canal irrigated area (‘000 hectare)			Remarks
	1970	1990	2018	
• Bist Doab (Hoshaipur, Kapurthala, Jalandhar, Nawanshaher)	40	53	4	Decreased by 36 thousand hectares
• Eastern Malwa (Ropar, SAS Nagar)	2	2	7	Increased by 5 thousand hectares
• Central Malwa (Ludhiana, Barnala, Sangrur, Patiala, Fatehgarh Sahib)	224	186	138	Decreased by 86 thousand hectares (40%)
• South-Western Malwa (Ferozepur, Moga, Fardikot, Muktsar, Bhatinda, Mansa)	786	975	976	Increased by 190 thousand hectares (25%)
• Upper Bari Doab (Gurdaspur Amritsar, TarnTaran)	249	276	51	Decreased by 198 thousand hectares (80%)
• Punjab State	1,301	1,492	1,176	Decreased by 125 thousand hectares (10%)

829

830 The share of reduction in canal irrigated area is highest in Central Malwa (~23%) of
831 which Sirhind Canal Tract is a part. On the contrary, the canal water supplies have increased
832 in southwestern districts. This has led to water-logging issues in the south-western region
833 as farmers avoid utilizing groundwater for irrigation, owing to its poor quality. In the central
834 region, farmers have become highly dependent on groundwater, and many of the canal
835 water courses are no longer in use or have disappeared. The canals and distributaries in the
836 region have merely become conveyance canal to transport water from Bhakra dam to
837 southwestern districts and neighbouring states. In an article by Singh (2018), out of the total
838 14.54 MAF of allocated river water, Punjab is using eight MAF less water, while the share to
839 the neighbouring states has remained constant. Also, due to a lack of repair and
840 maintenance, the carrying capacity of the canal network has decreased over time. A decline
841 in the area under canal irrigation could reduce the water flow in the rivers. A concrete canal
842 water redistribution policy and concentrated efforts to repair the network and restore field
843 distributaries can significantly sustain the declining groundwater table. The recent findings
844 by Kaur (2019) for the region point towards increasing rainfall trends by 2050 under
845 Representative Concentrated Pathways 4.5, which may also contribute to the increase in
846 canal water supplies.

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